



ELSEVIER

Contents lists available at ScienceDirect

Materials Science in Semiconductor Processing

journal homepage: www.elsevier.com/locate/msspPhotocatalytic, Fenton and photo-Fenton degradation of RhB over Z-scheme g-C₃N₄/LaFeO₃ heterojunction photocatalystsYongchun Ye^{a,b}, Hua Yang^{a,b,*}, Xiangxian Wang^b, Wangjun Feng^b^a State Key Laboratory of Advanced Processing and Recycling of Non-ferrous Metals, Lanzhou University of Technology, Lanzhou 730050, China^b School of Science, Lanzhou University of Technology, Lanzhou 730050, China

ARTICLE INFO

Keywords:

LaFeO₃ nanoparticles
g-C₃N₄ nanosheets
g-C₃N₄/LaFeO₃ heterojunction
Catalytic activity
Photo-Fenton mechanism

ABSTRACT

Z-scheme g-C₃N₄/LaFeO₃ heterojunction photocatalysts were prepared by calcining the uniformly mixed g-C₃N₄ nanosheets and LaFeO₃ nanoparticles at 300 °C for 1 h. The as-prepared g-C₃N₄/LaFeO₃ composites were systematically characterized by XRD, SEM, TEM, XPS, UV–vis DRS, PL spectroscopy, EIS and photocurrent response. It is demonstrated that LaFeO₃ nanoparticles (average size: ca. 35 nm) are uniformly assembled onto the surface of g-C₃N₄ nanosheets through chemical bonding, leading to the formation of g-C₃N₄/LaFeO₃ heterojunctions. The photocatalytic, Fenton and photo-Fenton performances of the samples were investigated by degrading RhB using simulated sunlight as the light source. Compared to bare LaFeO₃ and g-C₃N₄, the g-C₃N₄/LaFeO₃ composites exhibit significantly enhanced photocatalytic and photo-Fenton degradation of the dye, and the maximum activity is observed for the 15%g-C₃N₄/LaFeO₃ composite. Moreover, the photo-Fenton performance of the composites is much higher than the photocatalytic or Fenton performances. The much improved photo-fenton activity is ascribed to the efficient separation of photogenerated electron-hole pairs due to the Z-scheme electron transfer and the synergistic effect between LaFeO₃ and g-C₃N₄. •OH radicals were examined by PL spectroscopy using TPA as a probe molecule, clearly revealing the •OH generation, and moreover, the •OH yield and the dye degradation rate show a similar variation between different cases of the catalysis process. This suggested that •OH is the dominant reactive species causing the dye degradation. The underlying photo-Fenton mechanism of the composites is discussed.

1. Introduction

In recent decades, perovskite-type oxides have been received great attention due to their potential application to deal with the two urgent problems of energy shortage and environmental pollution [1–6]. Perovskite-type oxides have the general formula ABO₃ (where A is a rare-earth or alkaline-earth metal element and B is a transition metal element). Due to their unique crystal structures and strong electron-electron interactions, perovskite-type oxides take on a rich variety of physicochemical properties such as catalytic, electrical, magnetic and optical properties. In particular, perovskite-type oxides generally have a relatively narrow bandgap and can efficiently absorb visible light, and moreover they are advantaged by good chemical stability and flexible synthesis. These make perovskite-type oxides attractive as efficiently visible-light-responsive photocatalysts for the degradation of organic pollutants and hydrogen evolution from water splitting. However, the shortcoming of most of the perovskite-type oxides is the high recombination rate of photogenerated electron-hole (e⁻-h⁺) pairs. Therefore, it is indispensable to reduce the unfavorable electron-hole

recombination process for enhancing the photocatalytic activity. Numerous strategies have been developed to improve the photocatalytic activity of perovskite-type oxides, such as doping with impurity elements [7,8], construction of heterojunction composites [9,10], decoration with noble metals [11,12] and creation of lattice defects [13,14].

It is well known that, among various wastewater treatment technologies, advanced oxidation processes (AOPs) (such as ozone-based oxidation, electrochemical oxidation, sonochemical oxidation, photocatalytic oxidation and Fenton/photo-Fenton oxidation) have been considered as a high-efficiency and low-cost technology to eliminate organic pollutants [15–20]. As one of the AOPs, Fenton or Fenton-like oxidation technology is particularly attractive, which is mainly based on the generation and utilization of hydroxyl (•OH) radical as the oxidizing agent. The •OH radical is a very powerful oxidizing agent, capable of indiscriminately oxidizing and degrading organic pollutants into harmless inorganic substances. However, the traditional homogeneous Fenton process has some drawbacks including narrow working pH range (pH 2–4), generation of iron sludge, catalyst loss in the effluent

* Corresponding author at: State Key Laboratory of Advanced Processing and Recycling of Non-ferrous Metals, Lanzhou University of Technology, Lanzhou 730050, China.
E-mail address: hyang@lut.cn (H. Yang).

<https://doi.org/10.1016/j.mssp.2018.03.033>

Received 12 February 2018; Received in revised form 20 March 2018; Accepted 25 March 2018

Available online 05 April 2018

1369-8001/ © 2018 Elsevier Ltd. All rights reserved.

and treatment of secondary pollution from acid or metal ions [21]. Heterogeneous Fenton oxidation has been developed to overcome these problems existing in the homogeneous Fenton process. Moreover, when the heterogeneous Fenton system is irradiated with ultraviolet (UV) or visible light, it brings about a synergistic effect between Fenton oxidation and photocatalysis, and as a result, the photo-Fenton process is much more effective for the oxidative degradation of organic pollutants than Fenton oxidation or photocatalysis. Recently, it was continuously reported that iron-based oxides exhibited great photo-Fenton performance, such as $\text{BaFe}_{12}\text{O}_{19}/\text{g-C}_3\text{N}_4$ [22], $\text{Fe}_3\text{O}_4/\text{C}/\text{Cu}_2\text{O}$ [23], and yolk-shell structured $\text{Fe}_3\text{O}_4/\text{void}/\text{TiO}_2$ [24]. More importantly, the formation of heterojunctions in these composites can facilitate the separation of photogenerated electron-hole pairs and induce the synergistic effect between iron-based oxide and another semiconductor. This offers a promising strategy for the further improvement of photo-Fenton performance. Among various perovskite-type oxides, LaFeO_3 has been increasingly studied not only due to its photocatalytic activity but also due to its Fenton or photo-Fenton performance [25–29]. Wang et al. [30] immobilized LaFeO_3 nanoparticles on carbon spheres to achieve an excellent photo-Fenton performance, much higher than the photocatalytic performance. In our previous study, we found that the photocatalytic performance of LaFeO_3 nanoparticles can be significantly enhanced by the decoration of Ag nanoparticles [31].

Graphite-like carbon nitride ($\text{g-C}_3\text{N}_4$), being a metal-free polymeric semiconductor with relatively small bandgap energy of ca. 2.8 eV, has been shown to be a very promising visible-light-responsive photocatalyst [32]. $\text{g-C}_3\text{N}_4$ has a sufficiently negative conduction band (CB) edge potential of ca. -1.2 V vs normal hydrogen electrode (NHE), endowing it with a strong photocatalytic reduction capability. Moreover, $\text{g-C}_3\text{N}_4$ exhibits a good thermal and chemical stability, and can be simply and cheaply prepared by heating carbon- and nitrogen-containing organic precursors such as urea, melamine, cyanamide, dicyandiamide and thiourea. Due to these outstanding properties, $\text{g-C}_3\text{N}_4$ has been extensively studied for the photocatalytic hydrogen generation from water splitting, reduction of carbon dioxide into useful fuels and removal of environmental pollutants [33–35]. Further, two-dimensional (2D) $\text{g-C}_3\text{N}_4$ nanosheets have large specific surface area and abundant surface groups, which makes $\text{g-C}_3\text{N}_4$ nanosheets especially suitable to incorporate with other semiconductors to form heterojunction composite photocatalysts with much enhanced photocatalytic performances [36–42]. Very recently, several research groups reported the construction of $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction composites and found that they showed enhanced photocatalytic degradation of dyes and photocatalytic water splitting into hydrogen [43–47]. However, there is no work concerned with the Fenton or photo-Fenton performances of $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction composites. In this work, we attempted to prepare Z-scheme $\text{g-C}_3\text{N}_4$ nanosheets/ LaFeO_3 nanoparticles heterojunction composites via a simple mixing–calcining method, and systematically and comparatively investigate their photocatalytic, Fenton and photo-Fenton performances toward the degradation of rhodamine B (RhB). The underlying photo-Fenton synergistic reaction mechanism was discussed.

2. Experimental

2.1. Preparation of $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction composites

The preparation of $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ composites was accomplished in three steps. The first step was to prepare LaFeO_3 nanoparticles via a polyacrylamide gel route. Stoichiometric amount of $\text{La}(\text{NO}_3)_3$ (0.0075 mol), $\text{Fe}(\text{NO}_3)_3$ (0.0075 mol), citric acid (0.0225 mol), glucose (20 g) and acrylamide (0.135 mol) were successively dissolved in 20 mL of dilute nitric acid solution (5 mL HNO_3 + 15 mL deionized water) under continuous stirring using a magnetic bar. The mixture solution was added with deionized water to make a total volume of 100 mL and heated at 80 °C for 4 h, during which time a gel was formed. The gel was

dried in a thermostat drier at 120 °C for 24 h, and then calcinated in a tubular furnace at 500 °C for 3 h, yielding final LaFeO_3 nanoparticles. The second step was to prepare $\text{g-C}_3\text{N}_4$ nanosheets by directly heating melamine in a semiclosed system. 5 g of melamine was loaded in a corundum boat with a cover, and submitted to calcination in a tube furnace at 520 °C for 4 h. The obtained product was ground and collected as $\text{g-C}_3\text{N}_4$ nanosheets. The third step was to prepare $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction composites. Stoichiometric amount of $\text{g-C}_3\text{N}_4$ and LaFeO_3 were ground together in an agate mortar for 1 h, and then submitted to heat treatment in a tubular furnace at 300 °C for 1 h. The product was collected as final $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction composite. By varying the mass fraction of $\text{g-C}_3\text{N}_4$ in the composite from 10% to 20%, several composite samples of 10% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$, 15% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ and 20% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ were prepared.

2.2. Sample characterization

The crystal phase of the as-prepared samples was determined by x-ray powder diffraction (XRD) using Cu K α radiation with wavelength of $\lambda = 0.15406$ nm on a D8 Advance x-ray diffractometer. A JSM-6701F field-emission scanning electron microscope (SEM) and a Tecnai G2 20 field-emission transmission electron microscope (TEM) were used to characterize the morphology, microstructure and chemical composition of the samples. The chemical states and composition of the samples were measured by x-ray photoelectron spectroscopy (XPS) on a Thermo Scientific Escalab 250Xi x-ray photoelectron spectrometer. The optical absorption and bandgap energy of the samples was investigated by ultraviolet-visible diffuse reflectance spectroscopy (UV-vis DRS) using BaSO_4 as the reference on a TU-1901 double-beam ultraviolet-visible spectrophotometer with an integrating sphere attachment. The photoluminescence (PL) spectra of the samples were measured on a LS-55 fluorescence spectrophotometer with the excitation wavelength of 315 nm.

2.3. Photoelectrochemical measurement

The separation and transfer behavior of photogenerated electrons and holes in the samples was investigated by photocurrent response and electrochemical impedance spectroscopy (EIS) on a CorrTest CST 350 electrochemical workstation. The photoelectrochemical measurements were carried out using a three-electrode cell configuration consisting of a working electrode, a platinum foil counter electrode and a standard calomel electrode (SCE) reference electrode. The working electrode was prepared by mixing together 15 mg of the photocatalyst (LaFeO_3 or 15% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$), 0.75 mg of carbon black and 0.75 mg of polyvinylidene fluoride (PVDF) using 1-methyl-2-pyrrolidone (NMP) as solvent to form uniform slurry. The slurry was uniformly coated onto fluorine-doped tin oxide (FTO) glass substrate with an effective area of 1×1 cm², followed by drying at 60 °C for 5 h in a thermostat drying oven. 0.1 mol L⁻¹ Na_2SO_4 aqueous solution was used as the electrolyte. A 200 W xenon lamp was used as the simulated sunlight source. The transient photocurrent response was measured at a bias potential of 0.2 V. The EIS measurement was carried out by the use of the sinusoidal voltage pulse with amplitude of 5 mV over a frequency range of 10^{-2} – 10^5 Hz.

2.4. Photocatalytic test

RhB in aqueous solution was chosen as the target organic pollutant to evaluate its photocatalytic degradation, Fenton degradation and photo-Fenton degradation behaviors over LaFeO_3 and $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ samples. A 200 W xenon lamp was used as the simulated sunlight source to evaluate the photocatalytic activity of the samples. In a typical photocatalytic experiment, 50 mg of the photocatalyst and 100 mL of RhB solution with concentration of 10 mg L⁻¹ were loaded into the photoreactor. The mixture was magnetically stirred for 30 min in the

dark to establish the adsorption/desorption equilibrium of RhB onto the photocatalyst surface. After that, the mixture was illuminated by a 200 W xenon lamp to initiate the photocatalytic reaction. To avoid the thermal effect on photocatalytic reaction, the photoreactor was cooled with a water-cooling system during the photocatalysis process to maintain the reaction solution at room temperature. At given intervals of 30 min, 2.5 mL of the reaction solution was taken out from the photoreactor to examine the concentration of RhB. The RhB concentration was determined by measuring the absorbance of the solution at a wavelength of $\lambda = 554$ nm on a UV–visible spectrophotometer. Before measurement, the photocatalyst particles were removed by centrifugation at 4000 rpm for 10 min. The degradation percentage of the dye is defined as $(C_0 - C_t)/C_0 \times 100\%$, where C_0 is the initial RhB concentration and C_t is the remaining RhB concentration after t min of photocatalysis. As a typical Fenton process for the degradation of RhB, 50 mg of the photocatalyst was loaded in 100 mL of RhB solution, followed by 30 min of magnetic stirring in the dark. Subsequently, 0.5 mL of H_2O_2 was added in the mixture to initiate the Fenton reaction. The whole process of the Fenton experiment was carried out in the dark. In the photo-Fenton process for the RhB degradation, 50 mg of the photocatalyst was loaded in 100 mL of RhB solution and then the mixture was magnetically stirred for 30 min in the dark. After that, the mixture was added with 0.5 mL of H_2O_2 and simultaneously illuminated by a 200 W xenon lamp to initiate the photo-Fenton reaction. During the Fenton and photo-Fenton process, the concentration change of RhB was monitored following the same procedure as described for the photocatalytic process.

2.5. Detection of hydroxyl

During the photocatalysis, Fenton and photo-Fenton process, the production of hydroxyl ($\cdot OH$) radicals over the photocatalysts was examined by PL spectroscopy using terephthalic acid (TPA) as a probe molecule [48]. A stoichiometric amount of TPA was dissolved in the NaOH aqueous solution with concentration of 1.0 mmol L^{-1} to make a 0.25 mmol L^{-1} TPA solution. To examine the $\cdot OH$ radicals produced during the photocatalysis process, 50 mg of the photocatalyst was loaded in 100 mL of the TPA solution. After magnetically stirred for 30 min in the dark, the mixture was irradiated by a 200 W xenon lamp. After reaction for 120 min, the reaction solution was sampled and centrifuged at 4000 rpm for 10 min to remove the photocatalyst. The upper clear solution in the centrifuge tube was used for the PL measurement at a LS-55 fluorescence spectrophotometer with an excitation wavelength of 315 nm. To examine the $\cdot OH$ radicals produced during the Fenton process, 50 mg of the photocatalyst was loaded in 100 mL of the TPA solution and then the mixture was submitted to magnetic stirring for 30 min in the dark. Subsequently, 0.5 mL of H_2O_2 was added in the mixture to initiate the Fenton reaction. The whole process of the Fenton experiment was carried out in the dark. After 120 min of the Fenton reaction, the reaction solution was sampled for the PL measurement. To examine the $\cdot OH$ radicals produced during the photo-Fenton process, 50 mg of the photocatalyst was loaded in 100 mL of the TPA solution and magnetically stirred for 30 min in the dark. After that, 0.5 mL of H_2O_2 was added in the mixture and simultaneously turn on a 200 W xenon lamp to initiate the photo-Fenton reaction. After Fenton reaction for 120 min, the reaction solution was sampled for the PL measurement.

3. Results and discussion

3.1. XRD analysis

Fig. 1 shows the XRD patterns of the as-prepared $LaFeO_3$, $g-C_3N_4$ and $g-C_3N_4/LaFeO_3$ composites, along with the standard XRD line pattern for $LaFeO_3$ orthorhombic structure (JCPDS#037-1493). It is seen that $g-C_3N_4$ displays two typical XRD diffraction peaks at 2θ of

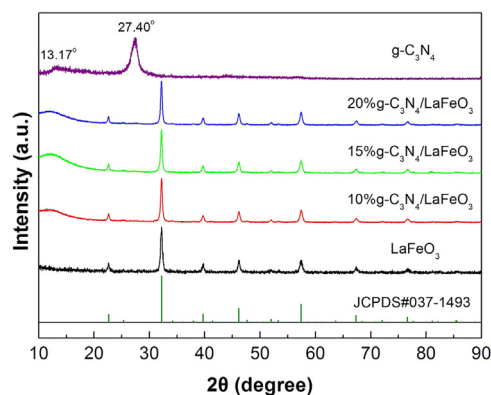


Fig. 1. XRD patterns of the as-prepared $LaFeO_3$, $g-C_3N_4$ and $g-C_3N_4/LaFeO_3$ composites, along with the standard XRD line pattern for $LaFeO_3$ orthorhombic structure (JCPDS#037-1493).

13.17° and 27.40° . The weak diffraction peak at 13.17° can be indexed to the (100) crystal plane of the graphite-like carbon nitride, which arises from the in-plane structural packing motif of tri-s-triazine units. The relatively strong diffraction peak at 27.40° refers to the (002) crystal plane of $g-C_3N_4$, which corresponds to the inter-layer stacking of conjugated aromatic system. This indicates the synthesis of $g-C_3N_4$ nanosheets using the present experimental method. For $LaFeO_3$, all the diffraction peaks can be perfectly indexed according to the JCPDS#037-1493 standard XRD line pattern, where the dominant diffraction peaks at $2\theta = 22.6^\circ, 32.2^\circ, 39.7^\circ, 46.2^\circ, 57.4^\circ$ and 67.4° correspond to reflections from (101), (121), (220), (202), (240) and (242) crystal planes of the $LaFeO_3$ orthorhombic phase, respectively. For the composites, all the diffraction peaks can be also indexed to the $LaFeO_3$ orthorhombic structure, indicating that there is no structural change in $LaFeO_3$. No obvious diffraction peaks of $g-C_3N_4$ are observed in the composites, which is due to several reasons: (1) $g-C_3N_4$ accounts for a relatively small fraction of the $g-C_3N_4/LaFeO_3$ composites, (2) the atomic scattering factor of light elements like C and N is relatively small, and (3) the destruction of long-range-ordered crystal structure for $g-C_3N_4$ nanosheets generally results in a low diffraction intensity. In addition, it is seen that the width of the XRD peaks of $LaFeO_3$ in the composites is very similar to that found for bare $LaFeO_3$, indicating that the grain size of $LaFeO_3$ particles undergoes almost no change when they are decorated onto $g-C_3N_4$ nanosheets. The average grain size (D) of $LaFeO_3$ particles can be quantitatively evaluated from the XRD peaks using Debye Scherrer formula [49]: $D = k\lambda/(\beta - \beta_0)\cos\theta$, where k is the shape factor ($k = 0.9$), λ is the wavelength of Cu K α radiation ($\lambda = 0.15406$ nm), θ is the diffraction angle, β is the full-width at half maximum (FWHM) of the XRD peaks of $LaFeO_3$, and β_0 is the FWHM of the XRD peaks of a standard reference material (Si). The grain size of $LaFeO_3$ particles estimated from the strongest diffraction peak of (121) is 32 nm.

3.2. SEM and TEM analysis

Fig. 2 shows the SEM image of a typical 15% $g-C_3N_4/LaFeO_3$ composite, revealing that $LaFeO_3$ particles are uniformly decorated onto two-dimensional $g-C_3N_4$ nanosheets. The $LaFeO_3$ particles are seen to present sphere-like or ellipsoid-like morphology and have a size distribution ranging from 20 to 60 nm. The average size of the particles is estimated to be 35 nm, which is very close to (only slightly higher than) their average grain size estimated from XRD data.

TEM investigation was also carried out to further elucidate the morphology and microstructure of $g-C_3N_4/LaFeO_3$ composites. Fig. 3(a) shows the TEM image of the 15% $g-C_3N_4/LaFeO_3$ composite. It is demonstrated that $g-C_3N_4$ exhibits a typical 2D sheet structure and $LaFeO_3$ particles are shaped like spheres or ellipsoids with an average size of

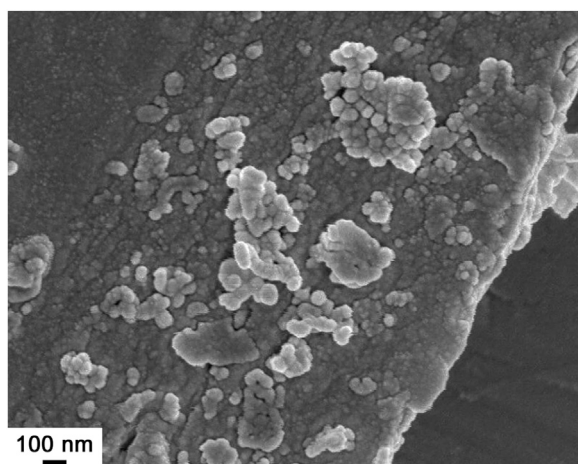


Fig. 2. SEM image of 15%g-C₃N₄/LaFeO₃.

35 nm. The LaFeO₃ particles are uniformly decorated onto the surface of g-C₃N₄ nanosheets almost without aggregation. The morphology and microstructure of the composites observed by TEM is basically in agreement with that observed from the SEM image. Fig. 3(b) shows the high resolution TEM (HRTEM) image of the 15%g-C₃N₄/LaFeO₃ composite, from which one can see that g-C₃N₄ has a typical amorphous lamellar structure and LaFeO₃ exhibits perfect crystal lattice fringes with no internal defects. The lattice fringes with the d-spacing of 0.395 nm correspond to the (101) crystal plane of LaFeO₃. The good assembly of LaFeO₃ particles onto the surface of g-C₃N₄ nanosheets leads to the formation of g-C₃N₄/LaFeO₃ heterojunction. Fig. 3(c)

shows the dark-field scanning TEM (DF-STEM) image of the 15%g-C₃N₄/LaFeO₃ composite, and Fig. 3(d)–(h) display the corresponding elemental mapping images of the region marked by yellow rectangle in Fig. 3(c). The bright particles present the uniform elemental distribution of La, Fe and O, and are determined to be LaFeO₃. The gray lamellar sheet displays the distribution of C and N elements, and is determined to be g-C₃N₄. The elemental mapping images further confirm that LaFeO₃ particles are uniformly assembled onto g-C₃N₄ nanosheets to form g-C₃N₄/LaFeO₃ heterojunction. The chemical composition of g-C₃N₄/LaFeO₃ composites is further investigated by the energy-dispersive x-ray spectroscopy (EDS). Fig. 3(i) shows the EDS spectrum of the 15%g-C₃N₄/LaFeO₃ composite, indicating the presence of La, Fe, O, C and N elements in the composite. The observed Cu signal could arise from the Cu microgrid that was used for supporting the sample. It is generally accepted that EDS is sensitive to heavy elements (e.g. La and Fe), but not to light elements (e.g. O, C and N), implying that the relative content of heavy elements can be precisely determined by EDS analysis. The atomic ratio of La/Fe is obtained as 1/1 from the EDS spectrum, which is perfectly consistent with the La/Fe atomic ratio in LaFeO₃ phase. However, from the EDS spectrum, it is difficult to determine the actual content of O in the LaFeO₃ phase as well as the g-C₃N₄ content in the g-C₃N₄/LaFeO₃ composite.

3.3. XPS analysis

XPS is an important surface analysis technique that can be used to analyze the chemical states and composition of the samples. The XPS analysis was performed on the 15%g-C₃N₄/LaFeO₃ composite, as shown in Fig. 4. Fig. 4(a) presents the XPS survey scan spectrum, revealing that the elements La, Fe, O, C and N are clearly included in the composite without other impurity elements. The observation of highly intensive C

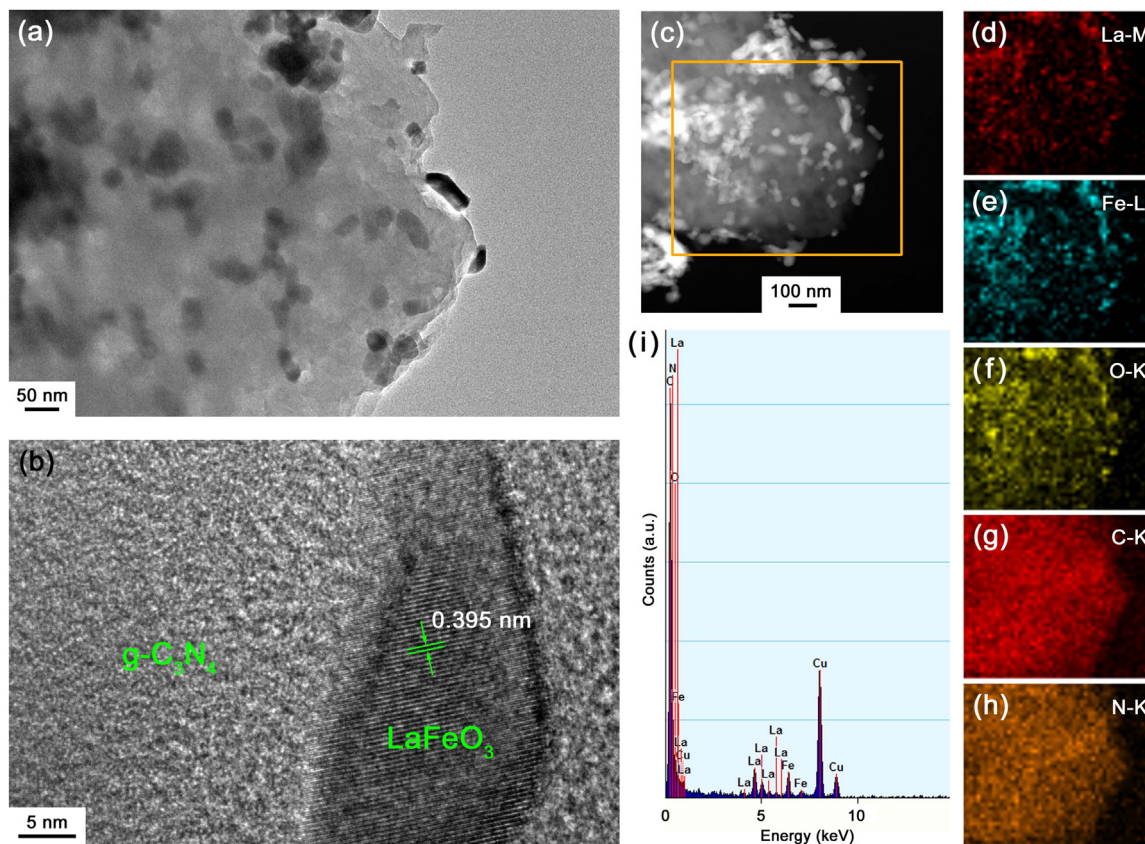


Fig. 3. (a) TEM image of 15%g-C₃N₄/LaFeO₃. (b) HRTEM image of 15%g-C₃N₄/LaFeO₃. (c) DF-STEM image of 15%g-C₃N₄/LaFeO₃. (d)–(h) The corresponding elemental mapping images of the region marked by yellow rectangle in (c). (i) EDS spectrum of 15%g-C₃N₄/LaFeO₃. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

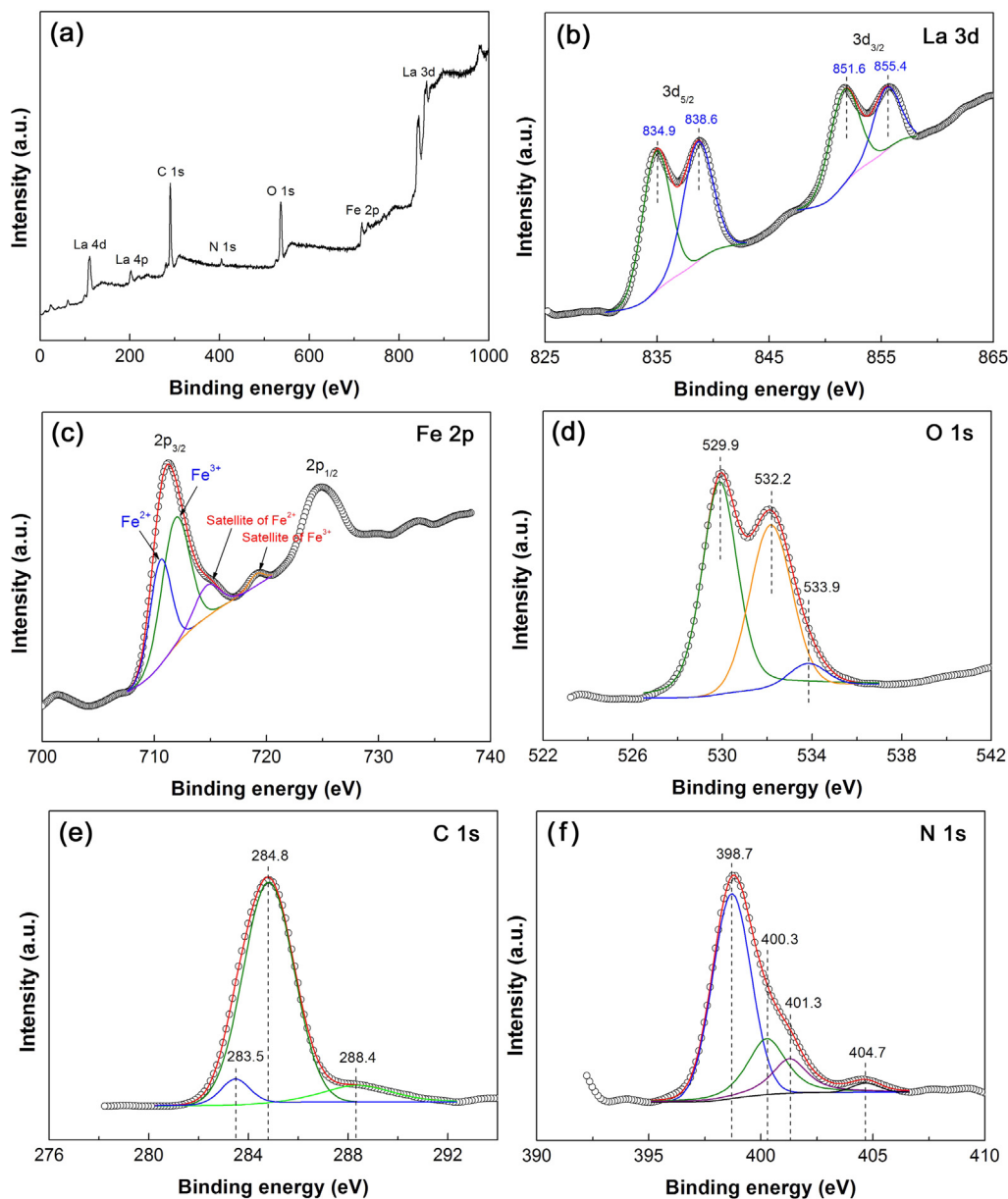


Fig. 4. XPS spectra of 15%g-C₃N₄/LaFeO₃. (a) XPS survey scan spectrum; (b) La 3d high-resolution XPS spectrum; (c) Fe 2p high-resolution XPS spectrum; (d) O 1s high-resolution XPS spectrum; (e) C 1s high-resolution XPS spectrum; (f) N 1s high-resolution XPS spectrum.

peak is due to the fact that adventitious carbon (C 1s at 284.8 eV) was used to calibrate the binding energy scale. Fig. 4(b)–(f) show the high-resolution XPS spectra of La 3d, Fe 2p, O 1s, C 1s and N 1s, respectively. The La 3d spectrum presents four peaks, of which the peaks at 834.9 and 851.6 eV are attributed to the binding energies of La 3d_{5/2} and La 3d_{3/2}, respectively, and the peaks at 838.6 and 855.4 eV are characterized as the corresponding satellites of La 3d_{5/2} and La 3d_{3/2} [50]. The appearance of the satellites on the La 3d levels is due to the spin orbit interaction and electron transfer from the 2p orbital of an oxygen ligand to an empty 4f orbital of lanthanum ions [50]. The La 3d spectrum indicates that lanthanum is in the form of La³⁺ oxidation state. On the Fe 2p spectrum, the Fe 2p_{3/2} signal can be deconvoluted into two peaks at 710.5 and 711.8 eV, which are assigned to Fe²⁺ and Fe³⁺, respectively [51]. Moreover, the Fe²⁺ peak is associated with a satellite at 714.2 eV and the Fe³⁺ peak is associated with a satellite at 719.0 eV. This clearly shows that there exist two oxidation states of iron (Fe²⁺ and Fe³⁺) in the LaFeO₃ particles. The existence of Fe²⁺ oxidation state is usually due to the generation of oxygen vacancies, as confirmed from

the O 1s XPS spectrum. The O 1s signal can be deconvoluted into three peaks at 529.9, 532.2 and 533.9 eV. The binding energy at 529.9 eV is assigned to the crystal lattice oxygen of LaFeO₃ [16,17]. The binding energy at 532.2 eV is attributed to chemisorbed hydroxyl species, which is closely associated with oxygen vacancies [17]. The weak peak at binding energy of 533.9 eV is ascribed to surface-adsorbed water molecules [18]. On the C 1s XPS spectrum, three peaks at 283.5, 284.8 and 288.4 eV are observed. The main peak at 284.8 eV is related to sp² C–C bonds of adventitious carbon used for the binding energy calibration [52,53]. The peak at 288.4 eV is attributed to the sp²-hybridized carbon in N-containing aromatic rings (N=C=N) [52,53]. The peak at 283.5 eV could be ascribed to the formation of lanthanum carbide or iron carbide species as a result of the interaction between La/Fe atoms with C atoms. It is noted that the C 1s binding energy for metal carbides is generally positioned at around 283.5 eV [54]. By the peak deconvolution analysis, the N 1s XPS spectrum is seen to consist of four peaks at 398.7, 400.3, 401.3 and 404.7 eV, which correspond to sp²-hybridized nitrogen in the form of C=N–C, tertiary nitrogen N–(C)₃ groups, amino

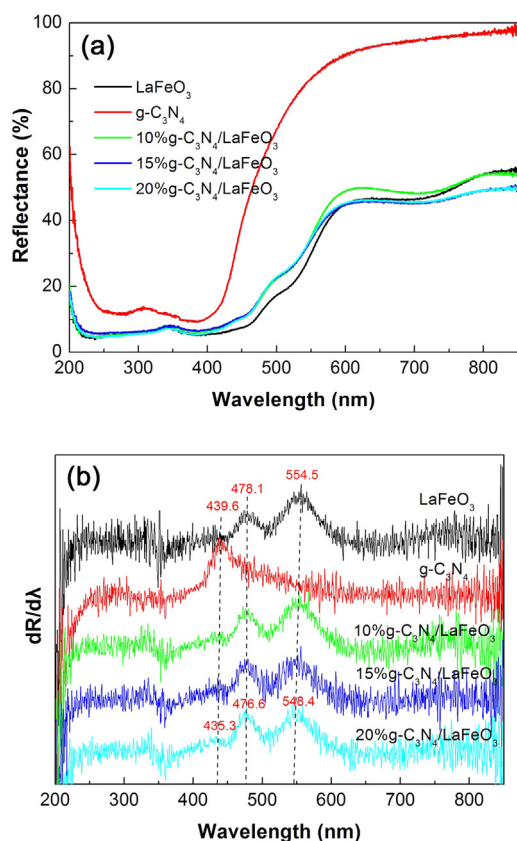


Fig. 5. (a) UV–vis DRS spectra of LaFeO₃, g-C₃N₄ and g-C₃N₄/LaFeO₃ composites. (b) The corresponding first derivative curves of the UV–vis DRS spectra.

functional groups with a hydrogen atom (C–N–H) and π -excitation, respectively [52,53]. This implies a typical characteristic feature of g-C₃N₄.

3.4. UV–vis DRS spectrum analysis

Fig. 5(a) shows the UV–vis DRS spectra of LaFeO₃, g-C₃N₄ and g-C₃N₄/LaFeO₃ composites, and Fig. 5(b) presents the corresponding first derivative curves of the UV–vis DRS spectra. The absorption edges of the samples associated with the electron transition from the valence band (VB) to the CB can be derived from the peaks on the first derivative spectra. It is seen that g-C₃N₄ nanosheets exhibit an absorption edge at 439.6 nm, whereas LaFeO₃ particles have two obvious absorption edges at 478.1 and 554.5 nm, which are related to the ligand to metal charge transfer (LMCT) and the metal to metal charge transfer (MMCT), respectively [31]. For g-C₃N₄/LaFeO₃ composites, the absorption edges of both LaFeO₃ and g-C₃N₄ exhibit a slight blue shift trend with the increase of g-C₃N₄ content in the composites. When the g-C₃N₄ content is increased up to 20%, that is, for the 20%g-C₃N₄/LaFeO₃ composite, the absorption edge of g-C₃N₄ blue-shifts to 435.3 nm, and the absorption edges of LaFeO₃ blue-shift to 476.6 (LMCT) and 548.4 nm (MMCT). From the absorption edges, the bandgap energies of the samples are obtained and given in Table 1. The bandgap energies of both LaFeO₃ and g-C₃N₄ in the composites exhibit a slight increase compared to those of bare LaFeO₃ particles and g-C₃N₄ nanosheets, which could be due to the chemical bonding between LaFeO₃ and g-C₃N₄.

3.5. PL spectrum analysis

To understand the photocatalytic activity of the composites, it is important to reveal the recombination behavior of photogenerated

Table 1
Bandgap energies of LaFeO₃, g-C₃N₄ and g-C₃N₄/LaFeO₃ composites.

Samples	Bandgap energies of LaFeO ₃ (eV)		Bandgap energy of g-C ₃ N ₄ (eV)
	LMCT	MMCT	
LaFeO ₃	2.59	2.24	–
g-C ₃ N ₄	–	–	2.82
10%g-C ₃ N ₄ /LaFeO ₃	2.59	2.25	2.83
15%g-C ₃ N ₄ /LaFeO ₃	2.60	2.26	2.84
20%g-C ₃ N ₄ /LaFeO ₃	2.60	2.26	2.85

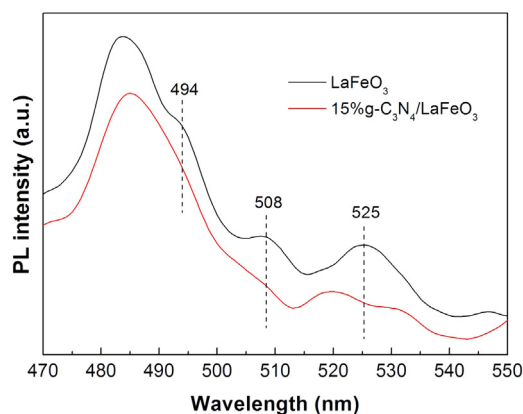


Fig. 6. PL spectra of LaFeO₃ and 15%g-C₃N₄/LaFeO₃.

electron-hole pairs in the composites. PL spectroscopy is an useful technique to evaluate the electron-hole pair recombination. Fig. 6 shows the PL spectra of LaFeO₃ and 15%g-C₃N₄/LaFeO₃. Several PL emission peaks are observed at around 494, 508 and 525 nm for bare LaFeO₃, and these PL signals arise due to the recombination of photo-generated electrons and holes. In contrast, the PL emission peaks from 15%g-C₃N₄/LaFeO₃ become obviously weakened, implying a decrease in the electron-hole recombination. As a result, more photogenerated electrons and holes in the g-C₃N₄/LaFeO₃ composites are available for participating in the photocatalytic reactions.

3.6. Photoelectrochemical analysis

The separation and transfer behavior of photogenerated electrons and holes in the composites were further investigated by photocurrent response and EIS. Fig. 7(a) shows the transient photocurrent responses of LaFeO₃ and 15%g-C₃N₄/LaFeO₃ measured for several on/off cycles of simulated sunlight illumination. When turning on the light, a photocurrent density of ca. 0.04 $\mu\text{A cm}^{-2}$ is observed for bare LaFeO₃, and ca. 0.09 $\mu\text{A cm}^{-2}$ for 15%g-C₃N₄/LaFeO₃. After turning off the light, the photocurrent density for both the samples drops to a very low level. This effect appears to be very reproducible when the light source is alternately switched on and off. The observed higher photocurrent density of 15%g-C₃N₄/LaFeO₃ than that of bare LaFeO₃ suggests an enhanced electron-hole pair separation in the composite. Fig. 7(b) shows the Nyquist plots of the EIS spectra for LaFeO₃ and 15%g-C₃N₄/LaFeO₃, both of which display a semicircle-like shape. It is generally accepted that a smaller diameter of the semicircle means a smaller charge-transfer resistance at the electrode/electrolyte interface. It is obvious that 15%g-C₃N₄/LaFeO₃ has a relatively smaller charge-transfer resistance due to the observation of smaller semicircle diameter as compared to bare LaFeO₃. This implies that the composite exhibits more efficient electron-hole separation and faster interface charge transfer under simulated sunlight illumination.

Based on the EIS measurements at different applied potentials (*V*) in the dark, the flat band potential (*V*_{FB}) of LaFeO₃ particles can be also

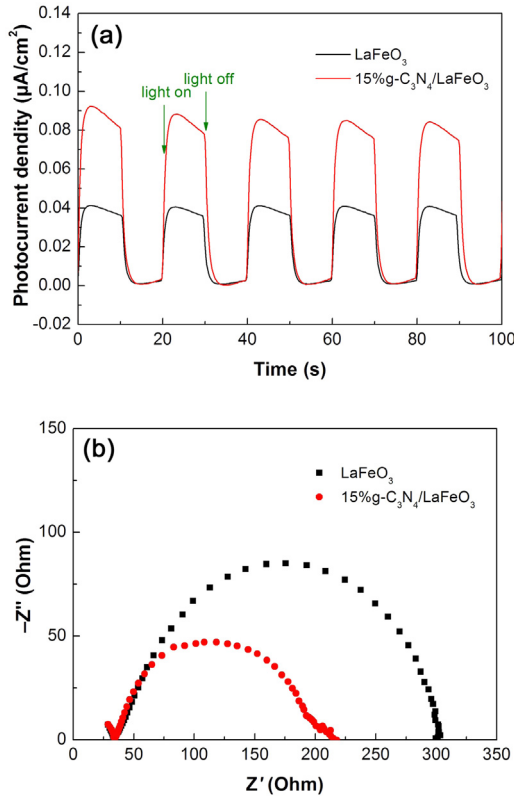


Fig. 7. (a) Transient photocurrent responses of LaFeO₃ and 15%g-C₃N₄/LaFeO₃. (b) Nyquist plots of the EIS spectra for LaFeO₃ and 15%g-C₃N₄/LaFeO₃.

derived according to the Mott-Schottky equation [55].

$$\frac{1}{C^2} = \left(\frac{2}{e\epsilon_r\epsilon_0 N_d A} \right) \left(V - V_{FB} - \frac{kT}{e} \right) \quad (1)$$

where C = space charge capacitance, e = electron charge, ϵ_r = relative permittivity, ϵ_0 = vacuum permittivity, N_d = majority carrier density, A = electrode surface area, k = Boltzmann constant, and T = absolute temperature. At room temperature, the temperature-dependent term kT/e can be negligible due to its small value. Thus, the flat band potential can be obtained by extrapolating the linear portion of the Mott-Schottky plot to the V axis. For each applied potential, the space charge capacitance is obtained from the capacitive zone of the Nyquist plots according to the following equation [56]

$$C = \frac{1}{2\pi f |Z_{im}|} \quad (2)$$

where $2\pi f$ is the angular frequency corresponding to the value of the imaginary component of the impedance Z_{im} . For a given frequency, the variation of $1/C^2$ against V is plotted as the Mott-Schottky plot.

Fig. 8 shows the Mott-Schottky plots at different frequencies (3000 and 5000 Hz) for LaFeO₃ particles. The two Mott-Schottky plots yield a similar flat band potential of + 1.74 V vs SCE. The negative slope of the Mott-Schottky plots implies that LaFeO₃ behaves as a p-type semiconductor. Assuming the gap between the flat band potential and the top edge of the VB is negligible for the p-type semiconductor, the VB potential of the p-type semiconductor is roughly equal to its flat band potential. As a result, the VB potential of LaFeO₃ is obtained as + 2.40 V vs NHE by converting the potential vs SCE reference electrode to the potential vs NHE according to $V(\text{NHE}) = V(\text{SCE}) + 0.059 \text{ pH} + 0.242$ (here pH = 7) [57]. By considering the MMCT bandgap energy for LaFeO₃ ($E_{g,\text{MMCT}} = 2.24 \text{ eV}$), the CB potential of LaFeO₃ is estimated to be + 0.16 V vs NHE.

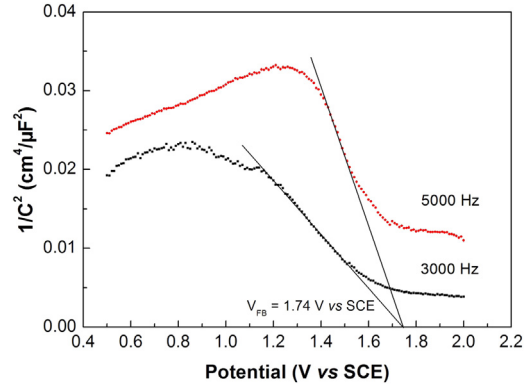


Fig. 8. Mott-Schottky plots of LaFeO₃ measured at 3000 and 5000 Hz.

3.7. Photocatalytic, Fenton and photo-Fenton performances

RhB, being one of the famous dyes and widely used as a colorant in paints, textiles and food products, is largely produced in the industrial dye wastewater. RhB is highly water-soluble, chemically stable, carcinogenic and non-biodegradable, and must be artificially removed. Therefore, RhB is generally chosen as a model substrate to evaluate the photocatalyst activity for the pollutant abatement. Fig. 9(a), (c) and (e) show the time-dependent photocatalytic, Fenton and photo-Fenton degradation of RhB over the samples, respectively. Prior to the dye degradation experiments, the dye adsorption was measured in the dark at 30 min of contact time and is obtained to be 4%–10% for all the samples. Moreover, the blank experiments demonstrate that RhB exhibits good stability during the photocatalytic, Fenton and photo-Fenton processes in the absence of the photocatalyst. In all cases of the photocatalytic, Fenton and photo-Fenton experiments, the degradation percentage of the dye after 120 min of reaction is given in Table 2. As seen from Fig. 9(a) and Table 2, the g-C₃N₄/LaFeO₃ composites exhibit a remarkably enhanced photocatalytic activity toward the dye degradation compared to bare LaFeO₃ nanoparticles and g-C₃N₄ nanosheets. With increasing the g-C₃N₄ content, the optimal photocatalytic activity is observed for the 15%g-C₃N₄/LaFeO₃ composite. We also measured the photocatalytic performance of the composites under UV irradiation and found a similar variation of the photocatalytic performance between the composites, implying that the dye sensitization via visible light absorption do not affect the conclusion of photocatalytic mechanism. From Fig. (c), it is observed that LaFeO₃ exhibits an obvious Fenton degradation of RhB, but g-C₃N₄ has almost no Fenton catalytic activity. The g-C₃N₄/LaFeO₃ composites show a decreased Fenton degradation of the dye compared to bare LaFeO₃, which is due to the introduction of g-C₃N₄ without Fenton catalytic activity, and moreover, the Fenton catalytic performance of the composites decreases monotonously with increasing the g-C₃N₄ content. For the photo-Fenton degradation of RhB as shown in Fig. 9(c), an enhanced photo-Fenton catalytic performance is observed for the g-C₃N₄/LaFeO₃ composites compared to bare LaFeO₃ and g-C₃N₄. Among the composites, the 15%g-C₃N₄/LaFeO₃ composite displays the highest photo-Fenton catalytic activity and the degradation percentage of the dye reaches 97.4% after 120 min of reaction. In addition, it is found that the photo-Fenton performance of the samples (particularly for LaFeO₃ and g-C₃N₄/LaFeO₃) is obviously higher than their photocatalytic performance and Fenton performance. This implies that the photo-Fenton catalysis is more efficient for the application in dye wastewater treatment. To further elucidate the catalytic activity of the samples in every case, the degradation kinetics of the dye is investigated. Fig. 9(b), (d) and (f) show the plots of $\ln(C_t/C_0)$ vs irradiation time t for the photocatalytic, Fenton and photo-Fenton degradation of RhB, respectively. It is seen that the dye degradation process conforms well to the first-order kinetic equation $\ln(C_t/C_0) = -k_{app}t$, where k_{app} is the apparent first-

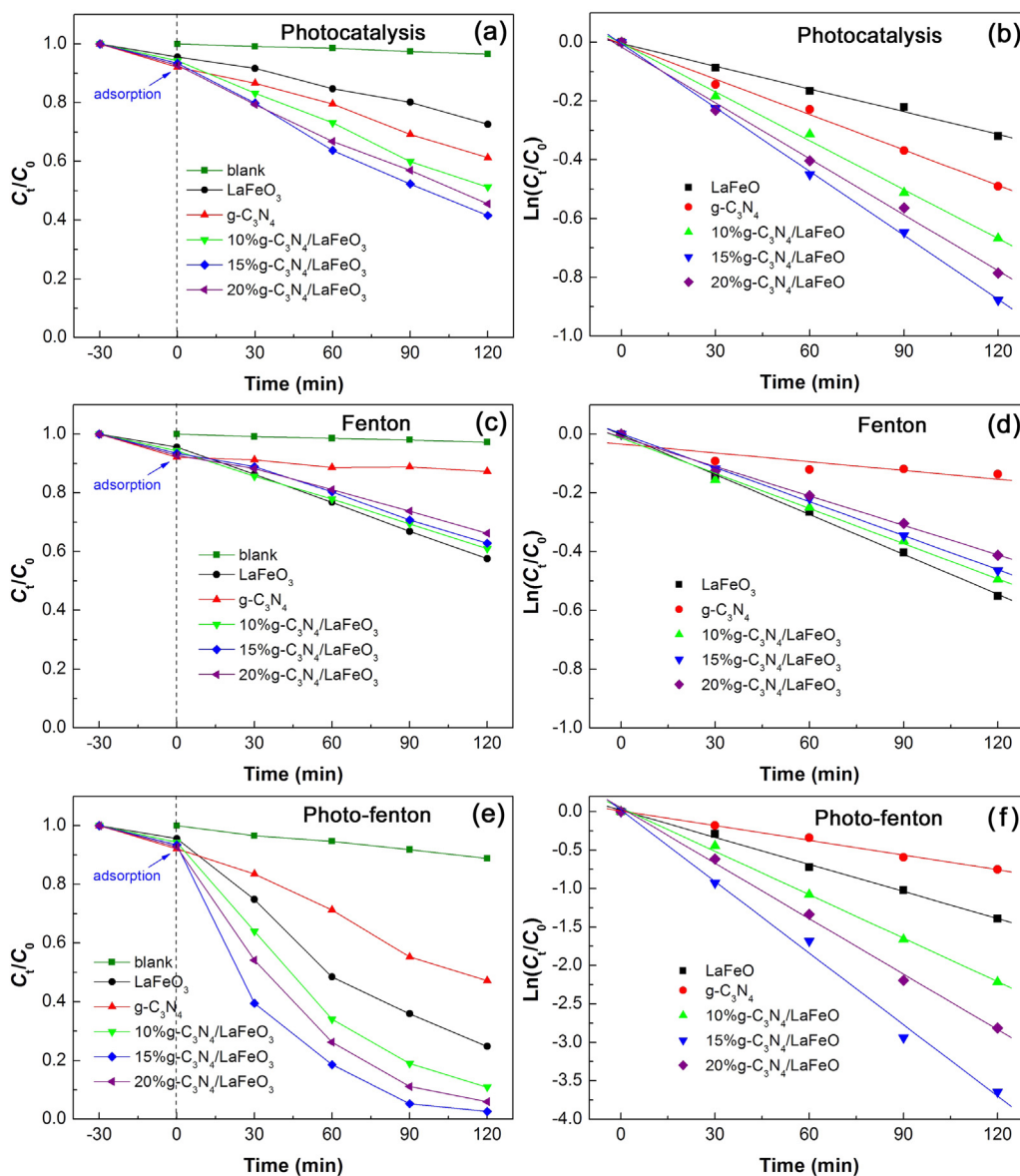


Fig. 9. (a), (c) and (e) Time-dependent photocatalytic, Fenton and photo-Fenton degradation of RhB over the samples, respectively. (b), (d) and (f) Plots of $\ln(C_t/C_0)$ vs irradiation time t for the photocatalytic, Fenton and photo-Fenton degradation of RhB, respectively.

order reaction rate constant (min^{-1}) [58]. From the slope of the regression lines, The rate constant k_{app} is obtained and given in Table 2. Based on the rate constant, it is concluded that the 15% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ composite has a photocatalytic activity about 2.8 and 1.8 times higher than those of bare LaFeO_3 and $\text{g-C}_3\text{N}_4$, respectively, and moreover the composite has a photo-Fenton catalytic activity about 2.6 and 4.9 times higher than those of bare LaFeO_3 and $\text{g-C}_3\text{N}_4$ respectively.

Table 2

Degradation percentage of the dye after 120 min of reaction and the apparent first-order reaction rate constant for the photocatalytic, Fenton and photo-Fenton processes over the samples.

Samples	Photocatalysis		Fenton		Photo-Fenton	
	Degradation (%)	K_{app} (min^{-1})	Degradation (%)	K_{app} (min^{-1})	Degradation (%)	K_{app} (min^{-1})
LaFeO_3	27.4	0.00258	42.4	0.00453	75.2	0.01174
$\text{g-C}_3\text{N}_4$	38.8	0.00402	12.7	0.00099	52.8	0.00638
10% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$	48.7	0.00555	39.1	0.00400	89.1	0.01883
15% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$	58.4	0.00726	37.2	0.00386	97.4	0.03100
20% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$	54.4	0.00635	33.8	0.00334	94.0	0.02403

The reusability is one of the important factors for the practical application of the catalyst in dye wastewater treatment. In the present study, the reusability of the 15% $\text{g-C}_3\text{N}_4/\text{LaFeO}_3$ composite for the photocatalytic, Fenton and photo-Fenton degradation of RhB was investigated by the recycling experiment. After the completion of the first cycle of the catalysis, the photocatalyst was collected by centrifugation, washed with deionized water and then dried in a thermostat drying

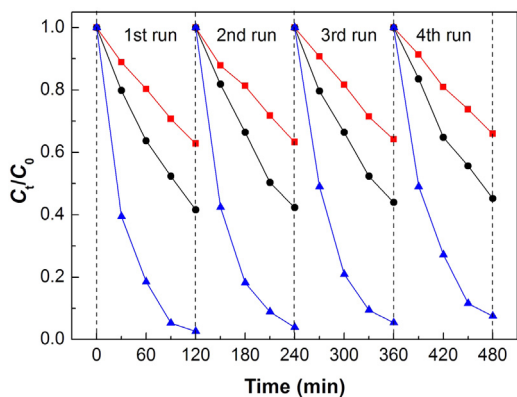


Fig. 10. Photocatalytic (black plot), Fenton (red plot) and photo-Fenton (blue plot) degradation of RhB over 15%g-C₃N₄/LaFeO₃ repeatedly used for four times. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.).

oven at 60 °C for 5 h. The recovered photocatalyst was used for the next cycle of the catalytic degradation of the dye. This process was repeated for three times. Fig. 10 shows the photocatalytic (black plot), Fenton (red plot) and photo-Fenton (blue plot) degradation of RhB during four catalytic cycles over the 15%g-C₃N₄/LaFeO₃ composite. At the 4th cycle of the photocatalysis, Fenton catalysis and photo-Fenton catalysis, the dye degradation after 120 min of reaction reaches 54.8%, 34.1% and 92.4%, respectively, compared to 58.4%, 37.2% and 97.4% at the 1st cycle of the corresponding catalytic degradation of the dye. This implies that the composite retains a good catalytic stability after recycling for four times.

The morphology, microstructure and crystal structure of the 15%g-C₃N₄/LaFeO₃ composite after four times of photocatalytic recycling were investigated by TEM and XRD. The TEM image shown in Fig. 11(a) reveals that LaFeO₃ particles are uniformly decorated onto two-dimensional g-C₃N₄ nanosheets. The HRTEM image given in Fig. 11(b) demonstrates that LaFeO₃ displays perfect crystal lattice fringes with the d-spacing of 0.279 nm, which correspond to the (121) facet of LaFeO₃. The XRD pattern presented in Fig. 11(c) shows that all the diffraction peaks can be perfectly indexed according to the LaFeO₃ orthorhombic phase (JCPDS#037-1493). The TEM and XRD investigation reveals the good assembly of LaFeO₃ particles onto g-C₃N₄ nanosheets and the microstructure and crystal structure of the composite undergoes almost no change after four times of photocatalytic recycling.

3.8. •OH radical analysis

In most of the photocatalysts including LaFeO₃, •OH radicals have been shown to be the dominant reactive species causing the dye degradation. In the present study, the production of •OH during the photocatalysis, Fenton and photo-Fenton process was investigated by PL spectroscopy using TPA as a probe molecule. It is noted that TPA can react with •OH to produce 2-hydroxyterephthalic acid (TAOH), which is an excellent fluorescent compound with a characteristic photoluminescence at around 429 nm. The intensity of the PL peak is related to the amount of produced •OH radicals. Fig. 12(a) shows the PL spectra of the TPA solution after 120 min of the photocatalysis, Fenton or photo-Fenton reaction over the samples, along with the blank experiment result. In all cases of the catalysis reaction, the TPA solution shows an obvious PL signal centered around 429 nm, whereas no PL signal is observed for blank TPA solution. This indicates that •OH radicals are

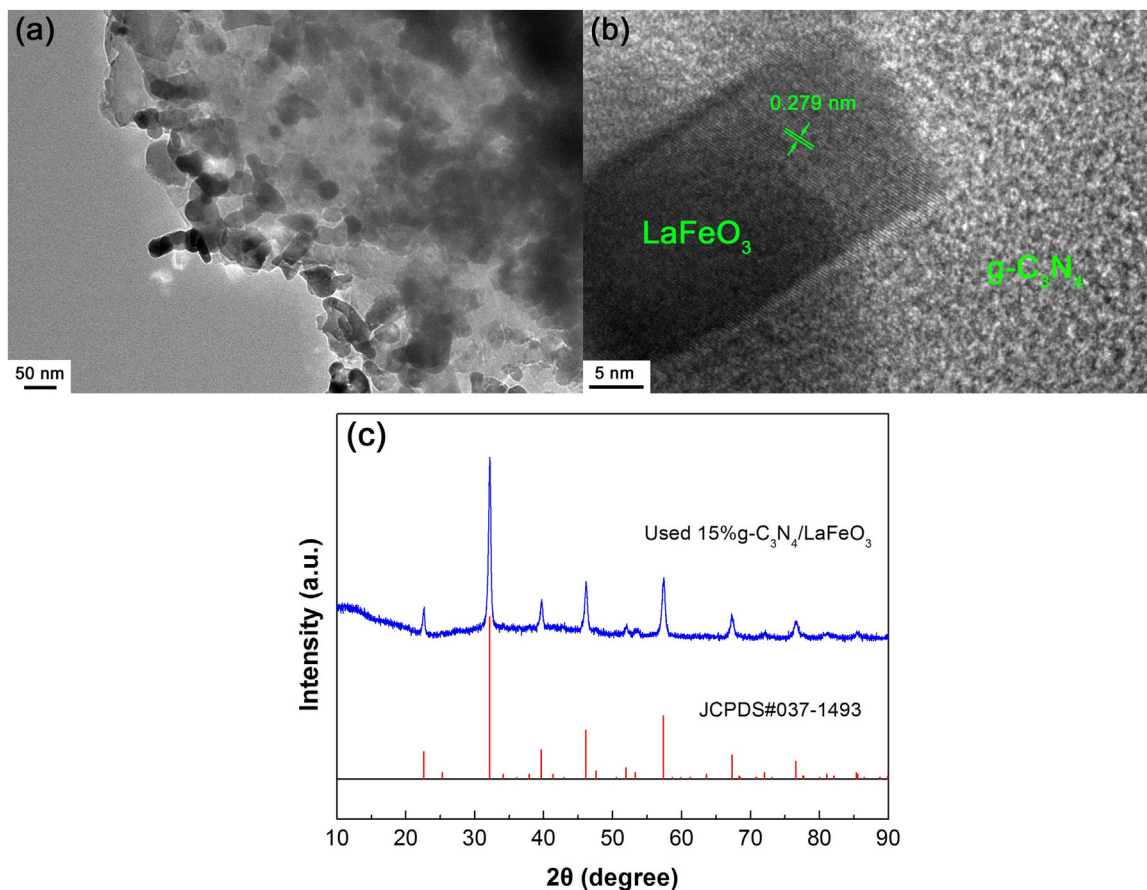


Fig. 11. TEM image (a), HRTEM image (b) and XRD pattern (c) of 15%g-C₃N₄/LaFeO₃ after four times of photocatalytic recycling.

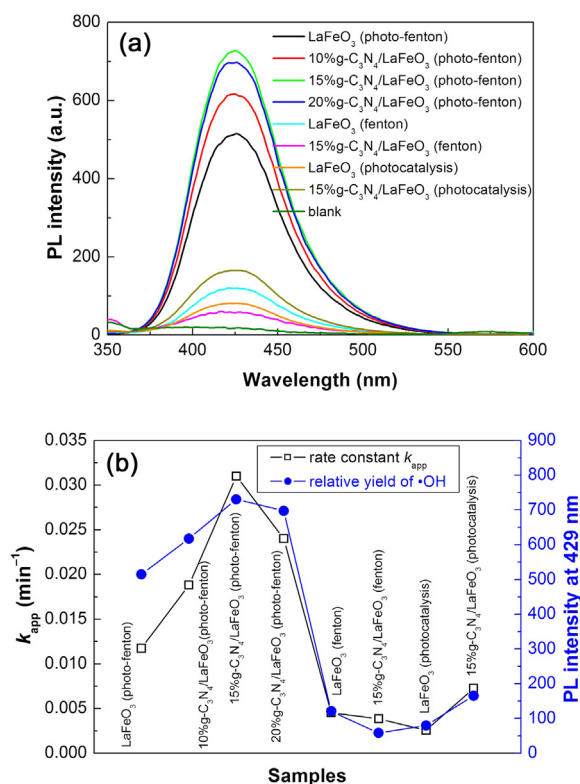


Fig. 12. (a) PL spectra of the TPA solution after 120 min of the photocatalysis, Fenton or photo-Fenton reaction over the samples, along with the blank experiment result. (b) Variation of the PL peak intensity and the rate constant k_{app} between different cases of the catalysis process.

produced in all cases of the catalysis process. The highest intensity of the PL peak is observed for the photo-Fenton reaction over the 15%g-C₃N₄/LaFeO₃ composite, implying the production of the largest amount of •OH radicals. In order to clearly elucidate the correlation of the •OH yield with the catalytic performance, the variation of the PL peak intensity (i.e. the •OH yield) and the rate constant k_{app} between different cases of the catalysis process are plotted in Fig. 12(b). It is seen that the variation of the rate constant basically agrees with the variation of the PL peak intensity, indicating that the catalytic performance is highly

dependent on the •OH yield and hence •OH radicals are confirmed to be the dominant reactive species responsible for the dye degradation. The existence of superoxide ($\bullet\text{O}_2^-$) can not be excluded in the present catalytic systems; however, $\bullet\text{O}_2^-$ radicals play only a minor or secondary role in the dye degradation [31].

3.9. Photo-Fenton mechanism

A possible photo-Fenton mechanism is proposed to elucidate the enhanced catalytic activity of the g-C₃N₄/LaFeO₃ heterojunction composites for the dye degradation, as shown in Fig. 13. The E_g of LaFeO₃ and g-C₃N₄ is 2.24 eV (here considering the typical MMCT E_g for LaFeO₃) and 2.82 eV, respectively. The CB/VB potentials of LaFeO₃ and g-C₃N₄ are obtained as + 0.16/ + 2.40 and - 1.18/ + 1.64 V vs NHE, respectively [31]. Due to their unique band edge positions, LaFeO₃ is expected to incorporate with g-C₃N₄ to form Z-scheme g-C₃N₄/LaFeO₃ heterojunction photocatalysts. Under irradiation of simulated sunlight, the photogenerated electrons in the CB of LaFeO₃ will transfer to g-C₃N₄ and recombine with the photogenerated holes in the VB of g-C₃N₄. The internal electric field at the interface of the g-C₃N₄/LaFeO₃ heterojunction acts as the driving force for the Z-scheme electron transfer, and moreover the formed chemical bonds between g-C₃N₄ and LaFeO₃ like C-La/Fe bonds as revealed by XPS and UV-vis DRS are also conducive to the electron transfer. Due to the Z-scheme electron transfer, more photogenerated holes in the VB of LaFeO₃ and photogenerated electrons in the CB of g-C₃N₄ are available for participating the photocatalytic reactions. As a result, the g-C₃N₄/LaFeO₃ composites exhibit enhanced photocatalytic performance compared to bare LaFeO₃ and g-C₃N₄. It is noted that the photogenerated holes in the VB of LaFeO₃ do not oxidize the dye molecules directly, but react with OH⁻ species to produce highly oxidative •OH radicals, which are confirmed to be the dominant reactive species causing the dye degradation. This •OH generation process is feasible from a thermodynamic point of view since the VB potential of LaFeO₃ (+ 2.40 V vs NHE) is positive to the redox potential of OH⁻/•OH (+ 1.99 V vs NHE) [52]. The CB potential of g-C₃N₄ (- 1.18 V vs NHE) is more negative than the redox potential of O₂/•O₂⁻ (- 0.33 V vs NHE) [48], implying that the photogenerated electrons in the CB of g-C₃N₄ can react with O₂ dissolved in the reaction solution to produce •O₂⁻ radicals. However, compared to •OH radicals, the effect of •O₂⁻ radicals on the dye degradation is very small and can be negligible. When adding H₂O₂ to the reaction solution, the photo-Fenton catalytic process is proceeded. Another important amount of •OH radicals are produced via the photo-Fenton process, i.e., the

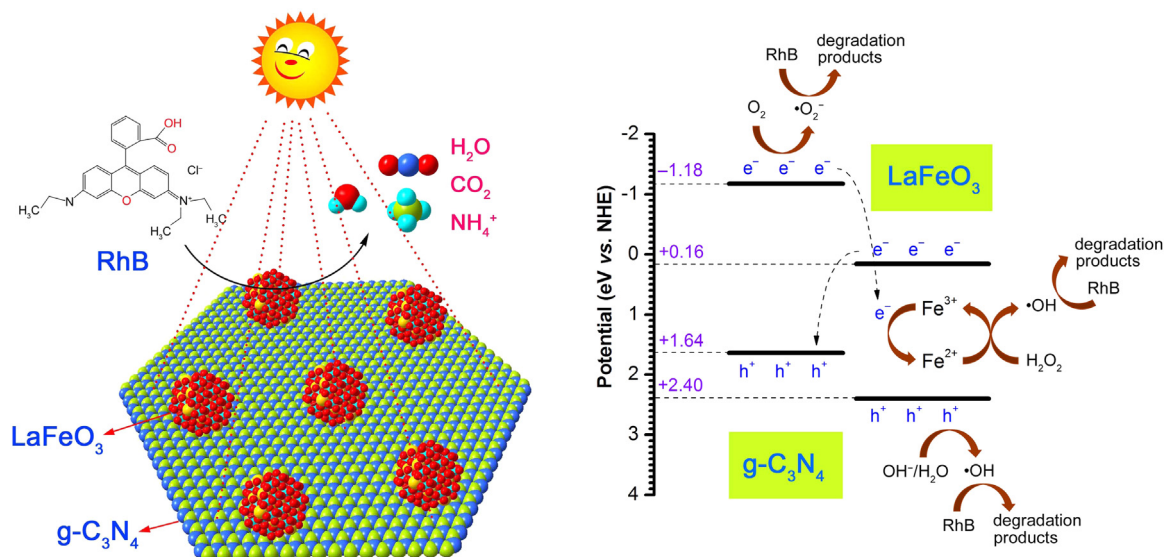
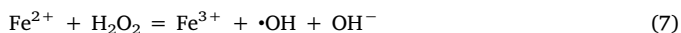
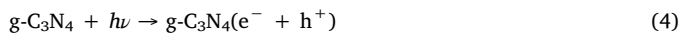
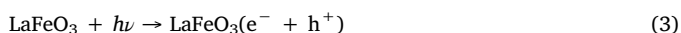


Fig. 13. Schematic illustration of the photo-Fenton mechanism for g-C₃N₄/LaFeO₃ composites.

reaction of Fe^{2+} with H_2O_2 [23,24]. Simultaneously, the photo-generated electrons in the CB of g- C_3N_4 cause the reduction of Fe^{3+} to Fe^{2+} , making the photo-Fenton process proceed continuously. Benefitting from this electron reduction, the photo-Fenton process for the g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ composites is more efficient than that for bare LaFeO_3 . Additional $\cdot\text{OH}$ radicals can be also generated by the direct UV photolysis of H_2O_2 . However, this is only a secondary pathway for the $\cdot\text{OH}$ generation because in the present study simulated sunlight was used as the light source, which contains only a small fraction of UV light. The main reactions involved in the photo-Fenton process can be briefly described by Eqs. (3)–(9). As seen from the above discussion, the creation of g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ heterojunctions is crucial for achieving excellent photocatalytic or photo-Fenton performance of the composites. An appropriate proportion of g- C_3N_4 and LaFeO_3 is required for the creation of the maximum amount of g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ heterojunctions in the composites. The optimum content of g- C_3N_4 in the composites is observed to be 15%, and as a consequence, the 15%g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ composite exhibits the highest photocatalytic or photo-Fenton activity toward the dye degradation.



4. Conclusions

LaFeO_3 nanoparticles with average size of ca. 35 nm were integrated with g- C_3N_4 nanosheets to form Z-scheme g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ heterojunction photocatalysts via a simple mixing–calcining method. The g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ composites demonstrate an efficient photogenerated electron-hole pair separation due to the Z-scheme electron transfer. As a result, the composites exhibit enhanced photocatalytic and photo-Fenton performances for the RhB degradation compared to bare LaFeO_3 and g- C_3N_4 . The maximum photocatalytic or photo-Fenton activity is observed for the 15%g- $\text{C}_3\text{N}_4/\text{LaFeO}_3$ composite. Moreover, the synergistic effect between LaFeO_3 and g- C_3N_4 makes the photo-Fenton performance of the composites much higher than the photocatalytic performance. $\cdot\text{OH}$ radicals are found to be generated in the photocatalytic, Fenton and photo-Fenton processes over the samples. The yield of $\cdot\text{OH}$ radicals varies consistently with the dye degradation rate between different cases of the catalysis process. It is concluded from the experimental results that $\cdot\text{OH}$ is the dominant reactive species responsible for the dye degradation.

Acknowledgements

This work was supported by the National Natural Science Foundation of China (Grant No. 51662027).

References

- [1] P. Kanhere, Z. Chen, *Molecules* 19 (2014) 19995–20022.
- [2] X.Y. Wu, X.Y. Wang, J. Li, G.K. Zhang, *J. Mater. Chem. A* 5 (2017) 23822–23830.
- [3] Y.X. Yan, H. Yang, X.X. Zhao, H.M. Zhang, J.L. Jiang, *J. Electron. Mater.* <<http://dx.doi.org/10.1007/s11664-018-6183-z>>.
- [4] N. Labhasetwar, G. Saravanan, S.K. Megarajan, N. Manwar, R. Khobragade, P. Doggali, F. Grasset, *Sci. Technol. Adv. Mater.* 16 (2015) 36002.
- [5] M. Zhou, H. Yang, T. Xian, R.S. Li, H.M. Zhang, X.X. Wang, *J. Hazard. Mater.* 289

- (2015) 149–157.
- [6] J. Shi, L. Guo, *Prog. Nat. Sci. Mater. Int.* 22 (2012) 592–615.
- [7] G. Iervolino, V. Vaiano, D. Sannino, L. Rizzo, V. Palma, *Appl. Catal. B-Environ.* 207 (2017) 182–194.
- [8] X.Q. Sun, Y.H. Xie, F.F. Wu, H.M. Chen, M.L. Lv, S. Ni, G. Liu, X.X. Xu, *Inorg. Chem.* 54 (2015) 7445–7453.
- [9] Y. Wu, Y.L. Wei, Q.Y. Guo, H. Xu, L. Gu, F.Y. Huang, D. Luo, Y.F. Huang, L.Q. Fan, J.H. Wu, *Sol. Energy Mater. Sol. Cells* 176 (2018) 230–238.
- [10] K. Nakamura, H. Mashiko, K. Yoshimatsu, A. Ohtomo, *Appl. Phys. Lett.* 108 (2016) 211605.
- [11] S.D. Li, Y. Zhao, C.Y. Wang, D.D. Li, K. Gao, *Mater. Lett.* 170 (2016) 122–125.
- [12] F. Wang, H. Yang, H.M. Zhang, J.L. Jiang, *J. Mater. Sci. -Mater. El* 29 (2018) 1304–1316.
- [13] H.Q. Tan, Z. Zhao, W.B. Zhu, E.N. Coker, B.S. Li, M. Zheng, W.X. Yu, H.Y. Fan, Z.C. Sun, *ACS Appl. Mater. Inter.* 6 (2014) 19184–19190.
- [14] L.J. Di, H. Yang, T. Xian, X.J. Chen, *Materials* 10 (2017) 1118.
- [15] H.W. Wang, X.Y. Li, Z.P. Hao, Y.J. Sun, Y.N. Wang, W.H. Li, Y.F. Tsang, *J. Environ. Manag.* 191 (2017) 244–251.
- [16] F.C. Moreira, R.A.R. Boaventura, E. Brillas, V.J.P. Vilar, *Appl. Catal. B-Environ.* 202 (2017) 217–261.
- [17] R. Behling, G. Chatel, S. Valange, *Ultrason. Sonochem.* 36 (2017) 27–35.
- [18] C.X. Zheng, H. Yang, Z.M. Cui, H.M. Zhang, X.X. Wang, *Nanoscale Res. Lett.* 12 (2017) 608.
- [19] T. Fan, Y.G. Li, H. Zhang, *Ind. Eng. Chem. Res.* 56 (2017) 11684–11696.
- [20] M.J. Lima, C.G. Silva, A.M.T. Silva, J.C.B. Lopes, M.M. Dias, J.L. Faria, *Chem. Eng. J.* 310 (2017) 342–351.
- [21] J. Herney-Ramirez, M.A. Vicente, L.M. Madeira, *Appl. Catal. B-Environ.* 98 (2010) 10–26.
- [22] H. Wang, Y. Xu, L. Jing, S. Huang, Y. Zhao, M. He, H. Xu, H. Li, *J. Alloy. Compd.* 710 (2017) 510–518.
- [23] F. Chai, K. Li, C. Song, X. Guo, *J. Colloid Interface Sci.* 475 (2016) 119–125.
- [24] D. Du, W. Shi, L. Wang, J. Zhang, *Appl. Catal. B-Environ.* 200 (2017) 484–492.
- [25] G. Iervolino, V. Vaiano, D. Sannino, L. Rizzo, P. Ciambelli, *Int. J. Hydrog. Energ.* 41 (2016) 959–966.
- [26] H.F. Shen, T. Xue, Y.M. Wang, G.Z. Cao, Y.J. Lu, G.L. Fang, *Mater. Res. Bull.* 84 (2016) 15–24.
- [27] X.C. Hao, Y.C. Zhang, *Mater. Lett.* 197 (2017) 120–122.
- [28] Y.L. Nie, L.L. Zhang, Y.Y. Li, C. Hu, *J. Hazard. Mater.* 294 (2015) 195–200.
- [29] V. Vaiano, G. Iervolino, D. Sannino, L. Rizzo, G. Sarno, P. Ciambelli, L.A. Isupova, *J. Environ. Eng.* 141 (2015) 04015038.
- [30] K. Wang, H. Niu, J. Chen, J. Song, C. Mao, S. Zhang, Y. Gao, *Appl. Surf. Sci.* 404 (2017) 138–145.
- [31] Y.C. Ye, H. Yang, R.S. Li, X.X. Wang, *J. Sol-Gel Sci. Technol.* 82 (2017) 509–518.
- [32] X. Wang, K. Maeda, A. Thomas, K. Takanabe, G. Xin, J.M. Carlsson, K. Domen, M. Antonietti, *Nat. Mater.* 8 (2009) 76–80.
- [33] S. Patnaik, S. Martha, S. Acharya, K.M. Parida, *Inorg. Chem. Front.* 3 (2016) 336–347.
- [34] F.L. Yang, F.F. Xia, J. Hu, C.Z. Zheng, J.H. Sun, H.B. Yi, *RSC Adv.* 8 (2018) 1899–1904.
- [35] J. Wen, J. Xie, X. Chen, X. Li, *Appl. Surf. Sci.* 391 (2017) 72–123.
- [36] Z.M. Cui, H. Yang, X.X. Zhao, *Mater. Sci. Eng. B* 229 (2018) 160–172.
- [37] J. Li, Y.C. Yin, E.Z. Liu, Y.N. Ma, J. Wan, J. Fan, X.Y. Hu, *J. Hazard. Mater.* 321 (2017) 183–192.
- [38] L. Zhou, W. Zhang, L. Chen, H. Deng, *J. Colloid Interface Sci.* 487 (2017) 410–417.
- [39] Z.X. Sun, H.Q. Wang, Z.B.A. Wu, L.Z. Wang, *Catal. Today* 300 (2018) 160–172.
- [40] S. Thaweesak, M. Lyu, P. Peerakiatkhajohn, T. Butburee, B. Luo, H. Chen, L. Wang, *Appl. Catal. B-Environ.* 202 (2017) 184–190.
- [41] Y. Wu, H. Wang, Y.M. Sun, T. Xiao, W.G. Tu, X.Z. Yuan, G.M. Zeng, S.Z. Li, J.W. Chew, *Appl. Catal. B-Environ.* 227 (2018) 530–540.
- [42] H. Wang, X.Z. Yuan, Y. Wu, G.M. Zeng, X.H. Chen, L.J. Leng, H. Li, *Appl. Catal. B-Environ.* 174 (2015) 445–454.
- [43] Q. Liang, J. Jin, C. Liu, S. Xu, Z. Li, *J. Alloy. Compd.* 709 (2017) 542–548.
- [44] K. Xu, H. Xu, G. Feng, J. Feng, *New J. Chem.* 41 (2017) 14602–14609.
- [45] S. Acharya, S. Mansingh, K.M. Parida, *Inorg. Chem. Front.* 4 (2017) 1022–1032.
- [46] K. Xu, J. Feng, *RSC Adv.* 7 (2017) 45369–45376.
- [47] Y. Wu, H. Wang, W. Tu, Y. Liu, Y.Z. Tan, X. Yuan, J.W. Chew, *J. Hazard. Mater.* 347 (2018) 412–422.
- [48] X. Zhao, H. Yang, Z. Cui, R. Li, W. Feng, *Mater. Technol.* 32 (2017) 870–880.
- [49] M.P. Klug, L.E. Alexander, *X-Ray Diffraction Procedure for Polycrystalline and Amorphous Materials*, Wiley, New York, 1974, p. 634.
- [50] A.J. Signorilli, R.G. Hayes, *Phys. Rev. B* 8 (1973) 81–86.
- [51] T. Yamashita, P. Hayes, *Appl. Surf. Sci.* 254 (2008) 2441–2449.
- [52] G.Q. Tan, L.N. She, T. Liu, C. Xu, H.J. Ren, A. Xia, *Appl. Catal. B-Environ.* 207 (2017) 120–133.
- [53] K.L. He, J. Xie, X.Y. Luo, J.Q. Wen, S. Ma, X. Li, Y.P. Fang, X.C. Zhang, *Chin. J. Catal.* 38 (2017) 240–252.
- [54] Q.D. Ling, S. Li, E.T. Kang, K.G. Neoh, B. Liu, W. Huang, *Appl. Surf. Sci.* 199 (2002) 74–82.
- [55] A. Fattah-alhosseini, *Arab. J. Chem.* 9 (2016) S1342–S1348.
- [56] F. Cardon, W.P. Gomes, *J. Phys. D-Appl. Phys.* 11 (1978) L63–L67.
- [57] F. Wang, H. Yang, Y.C. Zhang, *Mater. Sci. Semicond. Proc.* 73 (2018) 58–66.
- [58] I.K. Konstantinou, T.A. Albanis, *Appl. Catal. B-Environ.* 49 (2004) 1–14.